

Population genetics and history of the introduced fire ant, *Solenopsis invicta* Buren (Hymenoptera: Formicidae), in Australia

Michael T Henshaw,¹ Nicole Kunzmann,^{1,2} Cas Vanderwoude,³ Matthias Sanetra^{1,4} and Ross H Crozler^{1*}

¹School of Tropical Biology, James Cook University, Douglas, Qld 4611, Australia

²39 Girrawahen Dr, Gooseberrys Hill, WA 6075, Australia

³Fire Ant Control Centre, Queensland Department of Primary Industries, PO Box 1241, Oxley, Qld 4075, Australia

⁴Zoology and Evolutionary Biology, University of Konstanz, Universitätsstrasse 10, D 78457 Konstanz, Germany

Abstract The red imported fire ant, *Solenopsis invicta*, a damaging invasive pest, was discovered in February 2001 in Brisbane, Australia at two sites, Fisherman Islands and suburban Richlands-Wacol. Using four microsatellite loci and the protein marker *Gp-9*, we compared the two infestations with each other, and with potential source populations in North and South America to better understand the history of their introduction to Brisbane. Based on an analysis of molecular variance, as well as a maximum likelihood tree of colonies from the two Australian sites, we found that the two sites were genetically distinct and were almost certainly introduced separately. All of the colonies at Fisherman Islands were monogynous, headed by a single queen, while the Richlands-Wacol site had a mixture of single queen monogynous and multiple-queen polygynous colonies. However, the monogynous and polygynous colonies at the Richlands-Wacol site were not genetically distinct from each other, and probably constitute a single, mixed introduction. Based on allele frequencies at the microsatellite loci, and *Gp-9*, both Australian infestations were more similar to North American populations than to South American, though the Fisherman Islands infestation was intermediate, making it difficult to assign. Thus, there has been one introduction from either a North or South American monogynous population at Fisherman Islands, and one introduction from a mixed monogynous/polygynous North American population at Richlands-Wacol. These findings have implications for the control of the current infestations, as well as for the quarantine regulations necessary to prevent additional introductions to Australia.

Key words *Gp-9*, invasive species, microsatellites, red imported fire ant

INTRODUCTION

The red imported fire ant, *Solenopsis invicta*, is native to South America (Taber 2000), but has become a potential major exotic pest worldwide, as shown by expansion in the West Indies (Davis *et al.* 2001), New Zealand (Pascoe 2002), and Puerto Rico (Colloff & Collins 1996). It has been particularly destructive in North America. Since its introduction to North America in the 1930s (Tschinkel 1998), it has invaded over 120 million ha, and in Texas alone, costs annually over US\$1.2 billion in damage and for control (Lard *et al.* 2002). The invasion of *S. invicta* has devastating consequences for native invertebrate (Porter & Savignani 1990; Hu & Frank 1996; Forys *et al.* 2001; Stiles & Innes 2001; Wojcik *et al.* 2001) and vertebrate (Langfern *et al.* 1975; Allen *et al.* 1995, 1997a,b, 2001a,b; Lechner & Ribic 1996) communities.

In February 2001, *S. invicta* was found in Brisbane, Australia at two sites separated by approximately 30 km: Fisherman Islands, a port facility located at the mouth of the

Brisbane River, and suburban Richlands-Wacol, in south-west Brisbane (Matters & Vanderwoude 2001; McCubbin & Weiser 2002; Moloney & Vanderwoude 2002; Solley *et al.* 2002). The fire ant was estimated to have arrived only 3–5 years earlier, but had already occupied more than 40 000 ha (Vanderwoude *et al.* 2004). Based on the minimum and maximum climatic conditions endured by the fire ant, Vanderwoude *et al.* (2004) believed that it has the potential to invade and inhabit more than 95% of urban Australia. To prevent this outcome, an extensive 5-year program of control and eradication is underway (Vanderwoude *et al.* 2004). As part of such a program, as well as continuing efforts to prevent additional introductions, it is important to gain an understanding of the history of the fire ant's introduction to the Brisbane area.

Among the questions to be answered are whether multiple introductions have occurred and how they correspond to the known infestations. The Fisherman Islands and Richlands-Wacol infestations may have resulted from a single introduction that established itself at multiple sites, or there could have been separate introductions at each site. There may have even been multiple introductions at each site. The number of intro-

*Author to whom correspondence should be addressed (email: ross.crozler@jcu.edu.au)

ductions and how they have become established at different sites will have implications for quarantine measures, as well as for an understanding of the spread of the current infestation and how best to control it.

Knowledge of the epidemiology of the current infestations after they have become established is also important. In particular, there are two distinct social forms of *S. invicta* which have different population densities and modes of colony foundation that can affect their spread. The monogynous form has a single queen and effectively excludes foreign queens and workers from the colony. The polygynous form has many queens, and lacks effective colony discrimination (Glancey *et al.* 1973; Fletcher *et al.* 1980; Bhattachar & Vinson 1987; Tschinkel 1998). This allows polygynous colonies to exist at much higher densities than monogynous colonies (Mucron & Porter 1996), increasing their destructive potential. However, the monogynous form poses a greater challenge to effective monitoring because its colonies are founded by queens after a dispersal flight that can carry them several kilometres away (Markin *et al.* 1971; Tschinkel 1998). In contrast, polygynous colonies are dispersed through colony fission over short distances and adoption of new queens by other colonies (Glancey & Lafgren 1986). Previous work has suggested that the Fisherman Islands site is solely monogynous and that the Richlands-Wacol site has a mixture of monogynous and polygynous colonies.

A co-dominant mendelian locus, *General protein-9* (*Gp-9*), is strongly correlated with social form in *S. invicta* (Keller & Ross 1999), and can be used to characterise the social form of the Australian colonies. In polygynous colonies all of the queens are *Gp-9^{aa}* because *Gp-9^{bb}* queens are executed by the workers, and *Gp-9^{bb}* is lethal (Ross *et al.* 1999). In monogynous colonies, the queen is always *Gp-9^{aa}* (Goodisman *et al.* 2000). Thus all of her offspring, the workers and males, only have the B allele (mating between monogynous queens and polygynous males is believed to be extremely rare). There is also a cryptic b' allele in South American populations that functions as a b allele, but whose protein product behaves electrophoretically on starch gels as a B allele (Krieger & Ross 2002).

The source point for the incursion in Australia is also of quarantine interest. The fire ant has proven difficult to eradicate wherever it has established itself, and it is fortunate that it was discovered soon enough after its introduction in Brisbane that eradication is still possible. But the best hope for an *S. invicta* free Australia in the long term is effective quarantine. This will necessitate identifying the most likely sources of introduction.

In this study, we used highly polymorphic microsatellites as well as the *Gp-9* locus to characterise the population genetic structure of the Australian fire ant infestation and to confirm which social forms occur at each site. By characterising any differences between the two infestations, between the different social forms, and between the Australian infestations and North and South American populations, we determined how many introductions occurred, which social forms were introduced, and from where they were introduced.

MATERIALS AND METHODS

Fire ant collection

Between 7 October 2001 and 14 October 2001, we collected workers and males from 40 colonies at 20 sites in the Richlands-Wacol infestation, and 7 colonies at 4 sites in the Fisherman Islands infestation. We subsequently collected additional males from 13 colonies in the Richlands-Wacol infestation and 5 colonies in the Fisherman Islands infestation. All sampled colonies were putatively determined to be either monogynous or polygynous based on the size of the workers and the numbers of queens detected. The Queensland Department of Primary Industries also provided a single ant from each of 19 North American fire ant colonies (Florida, Georgia, Texas) and 7 South American colonies (Argentina) for comparison.

Extraction of genomic DNA

Genomic DNA was extracted using a commercially available salt precipitation protocol (Puregene DNA isolation kit, Gentra Systems). A minimum of 10 individuals, including both major and minor workers and males, were extracted from each nest. We used the whole ant for each extraction and the DNA was resuspended in 50 μ L of water for minor workers, 75 μ L for major workers, and 100 μ L for males.

General protein-9 genotyping

We genotyped 223 workers and 168 males at the *Gp-9* locus using previously published polymerase chain reaction (PCR) primers and protocols (Krieger & Ross 2002). Following PCR amplification, the products were subjected to a restriction fragment length polymorphism (RFLP) assay designed by Krieger (*pers. comm.* 2001) to distinguish between the *Gp-9^{aa}*, *Gp-9^{bb}*, *Gp-9^{b'}* genotypes. We digested 5 μ L of each sample with 1 μ L of the restriction enzyme *BsaAI* (5 units per μ L; New England Biolabs) overnight at 37°C in a 25 μ L reaction mixture. Following complete digestion, RFLPs were visualised on 4% agarose gels, stained with ethidium bromide, and photographed under UV light.

The RFLP assay was unable to distinguish between the b and b' alleles and we sequenced *Gp-9* from seven diploid males which were scored as having the *Gp-9^{bb}* genotype to determine whether any had the b' allele. Males are normally haploid, arising from unfertilised eggs, but diploid males can occur when fertilised, female-destined eggs are homozygous at the sex determining locus and develop as males instead. Diploid males frequently occur in populations that have reduced heterozygosity as a result of a bottleneck and they are a common occurrence in North American polygynous colonies. Sequences were obtained using the DYEnamic ET sequencing kit (Amersham Pharmacia) and visualised on an ABI 377 automated sequencer (Applied Biosystems).

Microsatellite PCR amplification

We genotyped the workers and males at four microsatellite loci: *sol 6*, *sol 11*, *sol 18*, and *sol 20* (Krieger & Keller 1997).

Microsatellite loci were amplified via PCR using 10 μ l. reactions containing 1 μ l. of PCR buffer (Promega 10x buffer containing 15 mM MgCl₂), 1 μ l. of 2.5 mM dNTPs (Amersham Pharmacia), 0.5 μ l. of 10 μ M fluorescently labelled forward primer, 0.5 μ l. of 10 μ M reverse primer, 0.1 μ l. of *Taq* polymerase (5 units per μ l.) and 2 μ l. of template DNA (Krieger & Keller 1997). Reactions were cycled through an initial denaturation at 92°C (2 min), followed by 35 cycles at 92°C (45 s), 55°C (45 s), and 72°C (1 min), and a final elongation step of 72°C (5 min). We visualised the PCR products on 5% denaturing polyacrylamide gels using the Gel-Scan 2000 electrophoresis system (Corbett Research).

Calculation of colony relatedness

We calculated relatedness among the workers in each colony based on their microsatellite genotypes according to the method of Queller and Goodnight (1989) using the computer program Relatedness 5.0 (Goodnight & Queller 1999). We estimated population allele frequencies in the Fisherman Islands and Richlands-Wacol populations separately and estimated standard errors by jackknifing over loci for colony estimates and over colonies for population estimates. We tested for differences in relatedness between populations and social forms using a one-way ANOVA and Fisher's Least Significant Difference (LSD) for *post hoc* tests.

Population structure and the number of introductions

To determine whether the Richlands-Wacol monogynous, polygynous, and Fisherman Islands colonies were introduced separately, we examined the partitioning of genetic variance, estimated from microsatellite allele frequencies, within and between infestation sites and social types using the Analysis of Molecular Variance (AMOVA) framework (Excoffier *et al.* 1992) implemented in the computer program Arlequin (Schneider *et al.* 2000). Significant partitioning of a large amount of the genetic variance between infestation sites or social forms would indicate differentiation, supporting the inference of multiple introductions.

Because the polygynous social form is only found at the Richlands-Wacol site, we also examined the partitioning of genetic variance between sites using only the monogynous colonies, and the partitioning of variance between social forms using only Richlands-Wacol colonies. The results did not change, and we only report the results for all colonies here.

To further assess the differentiation between infestation sites and social forms, we created both neighbour-joining and maximum likelihood trees of the colonies. We used the computer program *Distance* to compute Nei's D as a measure of genetic distance between colonies, *Neighbour* to create the neighbour-joining tree of colonies based on the genetic distances and *ConsML* to create the maximum likelihood tree of colonies. These programs are part of the *PhyIP* package of computer programs (Felsenstein 1993). The two trees were quantitatively the same and we only report the maximum likelihood tree.

Origins of the Australian invasion

To assess whether fire ants were introduced from North or South America, we also created neighbour-joining and maximum likelihood trees joining the Richlands-Wacol monogynous population, the Richlands-Wacol polygynous population, the Fisherman Islands monogynous population, the North American and South American samples provided by the Queensland Department of Primary Industries, and nine North American populations from Georgia and Louisiana whose allele frequencies were previously published (Ross *et al.* 1999). The two trees were quantitatively the same and we only report the maximum likelihood tree.

Finally, we compared the microsatellite allele frequencies in the Australian fire ants with the average allele frequencies in the nine North American populations (Ross *et al.* 1999) to determine whether the Australian and North American populations have similar subsets of the known alleles from South America (J. Keller pers. comm. 2002).

RESULTS

Social form

All of the workers and males at the Fisherman Islands site had *Gp-9^{aa}* and *Gp-9^b* genotypes, respectively (Table 1). The workers and males from the putatively monogynous colonies at the Richlands-Wacol site were also exclusively *Gp-9^{aa}* and *Gp-9^b*, respectively. The workers and males from the putatively polygynous colonies at the Richlands-Wacol site had a mixture of genotypes that included the *Gp-9^b* allele (Table 1). This allele is only found in polygynous colonies. Thus the Fisherman Islands site is strictly monogynous, while the Richlands-Wacol site is a mixture of monogynous and polygynous colonies.

None of the seven diploid males with *Gp-9^b* genotypes had the cryptic *b'* allele. This allele is found in some South American populations (Mescher *et al.* 2003), at a frequency of about 30%, but is not known in any North American populations. All of the sequenced males were from the Richlands-Wacol population and this result suggests that the Richlands-Wacol population is derived from North America. However, the 95% confidence interval for the frequency of the *b'* allele in the Richlands-Wacol population includes 0% to 41%, so a South American origin cannot be excluded on this evidence.

We also found that many of the males from the polygynous colonies were *Gp-9^{ab}* heterozygotes, indicating that they were diploid. Additional diploid males with *Gp-9^{aa}* and *Gp-9^{bb}* genotypes were detected because they were heterozygous at one or more of the microsatellite loci. Diploid males are a frequent occurrence in North American polygynous colonies, but not in monogynous colonies or in South America. This suggests that the polygynous colonies were introduced from North America. However, an alternative explanation is that they were introduced from South America and experienced a genetic bottleneck that reduced heterozygosity at the sex determining locus leading to increased production of diploid males.

Table 1 Frequency of *Gp-9* genotypes in *Solenopsis invicta* workers and males, within each social type

Genotype	Workers			Males		
	Fisherman Islands	Richlands-Wacol		Fisherman Islands	Richlands-Wacol	
	Monogynous (n = 34)	Monogynous (n = 53)	Polygynous (n = 94)	Monogynous (n = 56)	Monogynous (n = 65)	Polygynous (n = 47)
BB	1.0000	1.0000	0.3293	0.0000	0.0000	0.0551
Bb	0.0000	0.0000	0.6558	0.0000	0.0000	0.5057
bb	0.0000	0.0000	0.0149	0.0000	0.0000	0.1469
B				1.0000	1.0000	0.0638
b				0.0000	0.0000	0.1065

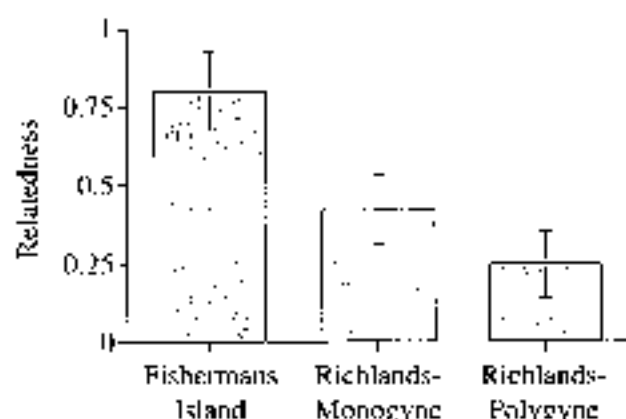


Fig. 1. Within-nest relatedness ($\pm 95\%$ confidence interval) in *Solenopsis invicta* determined for the Fisherman I and II site (all colonies monogynous) and for the monogynous and polygynous colonies separately at the Richlands-Wacol site, Brisbane.

Relatedness should be lower in polygynous colonies than in monogynous colonies because polygynous colonies have many queens. We found that relatedness in the putatively polygynous colonies was significantly lower than in either the monogynous Fisherman Islands, or monogynous Richlands-Wacol colonies as expected (Fig. 1). However, while workers at the Fisherman Islands colonies were approximately related as full sisters ($r = 0.75$), as would be expected under a single, once-mated queen, relatedness in the Richlands-Wacol monogynous colonies was significantly lower than in Fisherman Islands and was not representative of a single queen as expected.

Number of introductions

The AMOVA analysis revealed that 26% of the genetic variance in the Australian population was between the Richlands-Wacol and Fisherman Islands sites (Table 2), indicating that they are genetically distinct from each other and probably constitute separate introductions to Australia. Only 1% of the genetic variance was found between the polygynous and monogynous colonies in the Richlands-Wacol population (Table 2). This was not significantly different from zero, and they most likely constitute a single, mixed introduction from a locale where both social forms occur together.

Table 2 Results of the AMOVA analysis partitioning the genetic variation in *Solenopsis invicta* among and within collection sites, populations (Richlands-Wacol or Fisherman Islands), and social forms (monogynous or polygynous). The *P*-value estimates the probability of a more extreme variance component and *F*-statistic by chance alone

Variance component	Variance	% total	<i>F</i> -statistic	<i>P</i>
<i>Variance partitioned by collection site and social type</i>				
Between social types	0.01264	1.05	0.017	0.23
Within social type	0.23040	19.15	0.194	0.00065
Within collection sites	0.95537	79.77	0.202	<0.00001
<i>Variance partitioned by collection site and population</i>				
Between populations	0.35765	29.15	0.26182	0.00065
Within populations	0.11935	9.97	0.10800	<0.00001
Within collection sites	0.97500	80.85	0.84154	<0.00001

The maximum likelihood and neighbour-joining trees of the Australian colonies agree with the results of the AMOVA. The Fisherman Islands colonies cluster into a single group, and the monogynous and polygynous colonies in the Richlands-Wacol area all cluster together (Fig. 2). Thus, the two populations are distinct from each other and were probably introduced separately. The longest distance in the dendrogram is that from our South American sample in the nest, and according to the 'approximate confidence limits' provided by CONTML it is significantly longer than any other branch.

Source populations

The maximum likelihood tree of Australian, North American, and South American populations show the South American sample to be the most divergent, and clearly distinct from the Australian and North American populations (Fig. 3). The Richlands-Wacol population clusters closely with the North American populations. The Fisherman Islands population is more similar to the North American populations than to the South American sample, but it is intermediate, making it difficult to determine its origin with certainty.

Although the distance from the South American sample to the rest of the tree is significantly the largest distance, it may have been inflated by the effects of small sample size (Takezaki & Nei 1996). However, as for studies on the

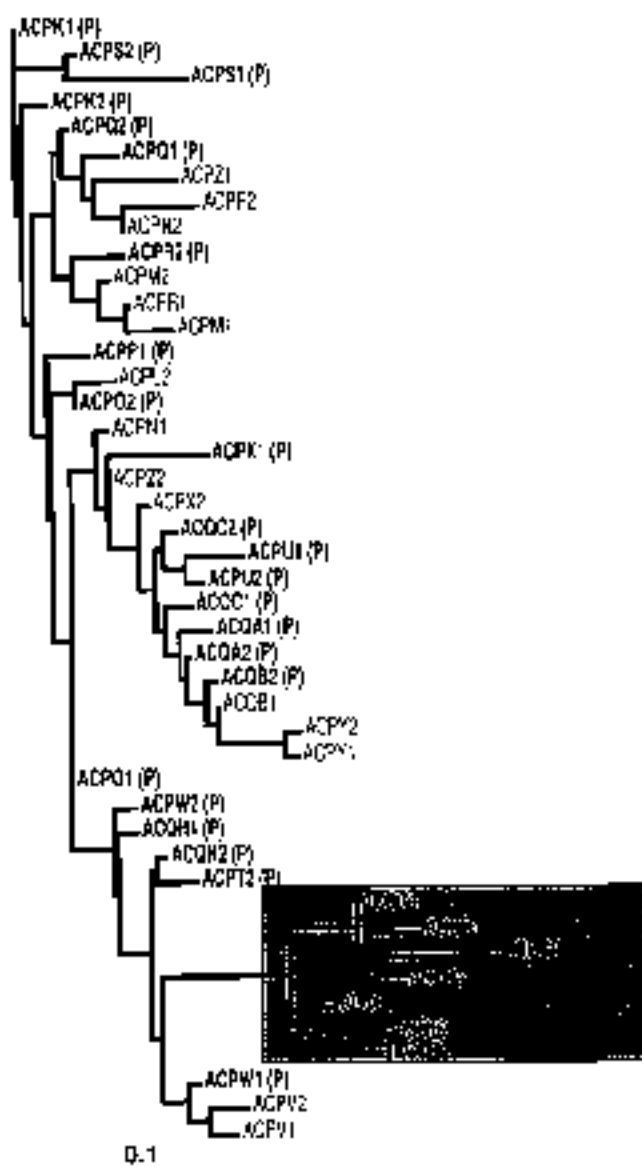


Fig. 2. Maximum likelihood tree of all *Solenopsis invicta* colonies sampled from Australia. Monogynous colonies are shown in grey and polygynous ones in black. The boxed colonies were those collected on Fisherman Islands. The scale gives expected accumulated variance (Felsenstein 1993).

Argentine ant, *Linepithema humile* Mayr (Tsutsui *et al.* 2001), we can also use the fact that introduced populations will have subsets of the alleles of the source areas. The Australian and North American populations had strikingly similar subsets of the known South American alleles at the four microsatellite loci (Table 3). There were a total of 58 known South American alleles at the four loci and 25 of these were also known from North America. Even though less than 50% of the known South American alleles were found in North America, 17 of the 21 alleles in the Australian populations were also in the North American subset of alleles.

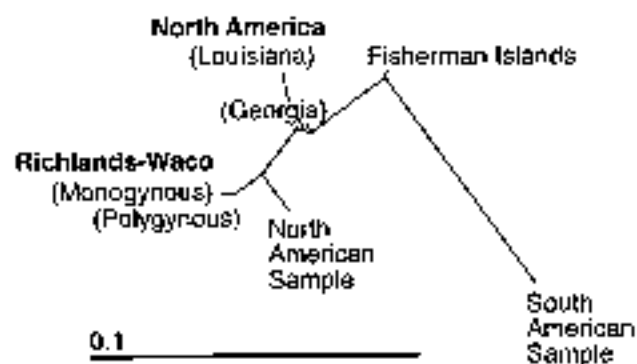


Fig. 3. Unrooted maximum likelihood tree of the three Australian populations of *Solenopsis invicta* together with published North American populations and small samples from North and South America.

DISCUSSION

Our results indicate that the two infestations at Fisherman Islands and Richlands-Waco are genetically distinct and have most probably been introduced separately. The Richlands-Waco site contains both monogynous and polygynous colonies, but there is no evidence that these are genetically distinct from each other at the microsatellite loci even though they are distinctly different at the *Gp-9* locus. This is most likely because the Richlands-Waco infestation was introduced from a mixed monogynous/polygynous population. Migration from monogynous populations into adjacent polygynous populations in North America has been shown to reduce genetic differences at neutral loci, such as microsatellites (Ross & Slinemaker 1993), while not affecting differences at *Gp-9* which are under selection (Grothman *et al.* 2000). Thus, there has been one introduction from a pure monogynous population at Fisherman Islands, and another introduction from a mixed monogynous/polygynous population in Richlands-Waco.

North America is the most likely source of both infestations. The evidence is particularly strong for the Richlands-Waco infestation. The polygynous colonies at Richlands-Waco have a high proportion of diploid males, a characteristic of North American polygynous colonies, but not of South American ones. Polygynous colonies in the Richlands-Waco infestation also appear to lack the *b*⁺ allele, which is found in some South American populations with a frequency of about 30% but is absent in North American populations. Finally, the Richlands-Waco infestation was highly genetically divergent from the South American sample, but extremely similar to North American populations as seen in the maximum likelihood tree. The Fisherman Islands infestation was also more similar to the North American populations than to the South American sample, but it was intermediate, making it difficult to assign. Mitochondrial DNA sequence information has been shown for Argentine ants to add to the picture obtainable from microsatellites (Tsutsui *et al.* 2001), and is being collected for

Table 3 Allele frequencies of four microsatellite loci in *Solenopsis invicta* in the Richlands-Wacol polygenous (RW-P), Richlands-Wacol monogenous (RW-M), and Fisherman Islands (FI) populations compared to the known alleles from South America (Argentina, I. Keller pers. comm. 2002), and the average population allele frequencies from nine previously published North American populations (Ress *et al.* 1999). Alleles in parentheses are not known from South America but were observed in either North American or Australian populations

Locus	Known South American alleles	RW-P (n = 23)	RW-M (n = 16)	FI (n = 7)	North America
Sei-6	87				
	95				0.001
	97				
	107				
	109	0.017		0.643	0.095
	111				0.002
	115	0.579	0.632	0.365	0.692
	116	0.408	0.364		0.270
	117				
	119				
	121	0.007	0.004		
	123				0.001
	125				
	127				
	129				
Sei-11	143	0.086	0.063	0.143	0.318
	145	0.091	0.142		0.060
	147	0.340	0.506		0.109
	149	0.086			0.001
	151	0.266	0.295	0.779	0.303
	153				
	155			0.014	0.209
	157				
	159				
	161				
	163				
	165			0.015	
	167				
	169				
	(171)				0.001
Sei-13	177				
	117				
	123				
	125	0.366	0.449	0.134	0.658
	127	0.472	0.419	0.266	0.141
	129	0.155	0.101		0.001
Sei-20	131				
	135				
	(139)	0.005	0.056		
	(141)				0.001
	120				
	122				
	124	0.271	0.499		0.146
	126	0.103	0.078	0.544	0.412
	128	0.120	0.161	0.226	0.168
	129				
	130			0.029	0.006
	131				
132				0.001	
134					
135				0.002	
138					
140					
141					
144				0.155	
146				0.001	
148	0.007				
150				0.006	
152	0.297	0.361		0.166	

Australian *S. invicta* by Prof. Jane Hughes (Curtin University, Brisbane).

The agreement between the Australian and North American allele size distributions indicates an origin of both invasive populations from North America. However, it is possible that the alleles found in North America were those of high frequency in South America and that a bottleneck in an Australian population originating from South America might lead to the same alleles entering Australia independently of the situation in North America. Populations of *S. invicta* outside the native range appear to be due to a single original introduction and this is reflected by their having similar subsets of the alleles to those seen in the native range (Tsuji et al. 2001).

Our results have several implications for the current plan to eradicate *S. invicta* from the Brisbane area and to prevent additional introductions to Australia. First of all, it is important that the two infestations are not the result of dispersal to multiple sites by a single introduction. Such a pattern of dispersal would indicate a high probability of other large undetected infestations in the south-east Queensland area which would jeopardise the current eradication program and dramatically increase the area to be monitored for successful eradication. However, it also suggests that additional quarantine effort may be necessary if the risk of future introductions is to be minimised, given that the current level has clearly been inadequate.

North America was almost certainly the source of one of the infestations and may also be the source of the other, suggesting that North America may be a more likely source than South America. This may be because of the greater volume of goods arriving from North America relative to South America (Anonymous 2003; Australian Bureau of Statistics 2003), or it could be because North American *S. invicta* occur at higher densities (Porter et al. 1997), making their accidental inclusion in shipped goods more likely (however, we acknowledge that there is no evidence that *S. invicta* were actually transported with shipped goods). Another possible explanation is that *S. invicta* arrived in Brisbane from another country or location where they have established but have not yet been detected.

These infestations in Brisbane have shown that, once introduced, *S. invicta* can remain undetected for years, quietly establishing themselves at high densities over large areas. Given the difficulty in eradicating fire ants once they are established, effective quarantine is the best means of minimising the risk of their establishment in Australia in the long term. However, effective quarantine for this and other pests should be supported by comprehensive surveillance programs, for example, bait and trap measures such as introduced at New Zealand entry points, to ensure new incursions are detected before economic and technical impediments (size of infested areas) reduce options for management or eradication of exotic pests.

ACKNOWLEDGEMENTS

This work includes material forming a thesis submitted in partial fulfilment of the requirements for a BSc (Hons) degree

by NK. We thank the Queensland Department of Primary Industries for financial support, Michael JB Krueger for the *Gp 9* assay, Sanford D Porter for discussions, Jo-Anne Holley for technical assistance, and Laurent Keller for the South American allele sizes. MTH held a James Cook University Postdoctoral Fellowship during this project. MS's stay in Australia was supported by a fellowship from the German Academy of Natural Scientists Leopoldina.

REFERENCES

- Allen CR, Demarus S & Lutz RS. 1997a. Effects of red imported fire ants on recruitment of white-tailed deer fawns. *Journal of Wildlife Management* **61**, 91–916.
- Allen CR, Forys EA, Rice KG & Wojcik DP. 2001a. Effects of the ant (Hymenoptera: Formicidae) on basking turtles and prevalence of fire ants on sea turtle nesting beaches in Florida. *Florida Entomologist* **84**, 251–253.
- Allen CR, Lutz RS & Demarus S. 1995. Red imported fire ant in goats: southern bobwhite populations. *Ecological Applications* **5**, 632–638.
- Allen CR, Lutz RS, Lockley J, Phillips SA & Demarus S. 2001b. The non-indigenous ant, *Solenopsis invicta*, reduces loggerhead snake and piglet inter-abundance. *Journal of Agricultural and Urban Entomology* **18**, 249–259.
- Allen CR, Rice KG, Wojcik DP & Percival HF. 1997b. Effects of red imported fire ant invasions on southern American alligators. *Journal of Herpetology* **31**, 318–321.
- Anonymous. 2003. *Looking to the Future: Part of Brisbane Corporation Annual Report 2002–2003*. Port of Brisbane Corporation, Brisbane, Australia.
- Australian Bureau of Statistics. 2003. *2000 International Merchandise Imports, Australia: Imports by State from Selected Countries and Country Groups, Customs Value (\$ million) January 1988 to October 2002*. Australian Bureau of Statistics, Canberra, Australia.
- Blaker AP & Yoccoz SB. 1987. Colony limits in *Solenopsis invicta* Buren. In: *Chemistry and Biology of Social Insects* (eds J Fahn & H Reinbold), pp. 399–400. Verlag J. Peperny, München, Germany.
- Collins AMA & Collins HC. 1996. Invasion and range expansion of red imported fire ant (Hymenoptera: Formicidae) in South America from 1918 to 1995. *Florida Entomologist* **79**, 240–251.
- Davis LR, Vander Meer RK & Porter SD. 2001. Red imported fire ants expand their range across the West Indies. *Florida Entomologist* **84**, 725–726.
- Excoffier L, Smouse PE & Quattro JM. 1992. Analysis of molecular variance inferred from metric distances among DNA haplotypes: application to human mitochondrial DNA restriction data. *Genetics* **131**, 479–491.
- Felsenstein J. 1985. *Phylogeny Inference Package, Version 3.5c*. Distributed by the author, Department of Genetics, University of Washington, Seattle, USA.
- Fletcher DM, Hlum MS, Whit TV & Temple N. 1990. Monogyny and polygyny in the fire ant, *Solenopsis invicta*. *Annals of the Entomological Society of America* **73**, 658–661.
- Frye EA, Quisoff A & Allen CR. 2001. Potential fire ant (Hymenoptera: Formicidae) impact on the endangered Schaus swallowtail (Lepidoptera: Papilionidae). *Florida Entomologist* **84**, 254–258.
- Gilmore BM, Craig CH, Stringer CE & Bishop PM. 1973. Multiple fertile queens in colonies of the imported fire ant, *Solenopsis invicta*. *Journal of the Georgia Entomological Society* **8**, 227–235.
- Gossney RM & Lofgren CS. 1988. Adoption of newly-mated queens: a mechanism for proliferation and perpetuation of polygynous red imported fire ants, *Solenopsis invicta* Buren. *Florida Entomologist* **71**, 581–587.
- Goodisman MAJ, Ross KG & Aronson MA. 2001. A formal assessment of the effects of gene flow and selection in the fire ant *Solenopsis invicta*. *Evolution* **54**, 600–610.

- Goodnight KF & Queller DC. 1995. Computer software for performing likelihood tests of pedigree relationship using genetic markers. *Molecular Ecology* 4, 1231–1234.
- Do GY & Frank JH. 1996. Effect of the red imported fire ant (*Hymenoptera: Formicidae*) on long-stabbing arthropods in Florida. *Environmental Entomology* 24, 1290–1296.
- Keller L & Ross KG. 1996. Major gene effects on phenotype and fitness: the relative roles of *Pgm-2* and *Cu-9* in introduced populations of the fire ant *Solenopsis invicta*. *Journal of Evolutionary Biology* 12, 672–680.
- Krieger MJB & Keller L. 1997. Polymorphism at *dimeric* and *monomeric* loci in fire ant *Solenopsis invicta* populations. *Molecular Ecology* 6, 997–999.
- Krieger MJB & Ross KG. 2002. Identification of a major gene regulating complex social behaviour. *Science* 295, 328–332.
- Lan C, Willis DB, Saha V & Robinson S. 2002. Economic assessments of red imported fire ant in Texas' urban and agricultural sectors. *Southern Economic Entomologist* 25 (Suppl. 1), 123–137.
- Lechner KA & Cook DU. 1996. Behavioural interactions between red imported fire ants (*Solenopsis invicta*) and three native species in south Texas. *Southern Economic Entomologist* 41, 124–128.
- Lofgren CS, Banks WA & Glenney BM. 1975. Biology and control of imported fire ants. *Annual Review of Entomology* 20, 1–30.
- Mason TE & Darter SD. 1996. Comparison of polygene and monogene red imported fire ants (*Hymenoptera: Formicidae*) population densities. *Annals of the Entomological Society of America* 89, 535–543.
- Moran GP, Diller JH, Hill SG, Blum MS & Hromova HR. 1971. Nuptial flight and flight longevity of the imported fire ant, *Solenopsis invicta* nomen novum (Hymenoptera: Formicidae). *Annals of the Georgia Entomological Society* 6, 145–156.
- McCubbin KI & Weiner JM. 2002. Fire ants in Australia: a new medical and ecological hazard. *Medical Journal of Australia* 176, 518–519.
- Mexler MC, Ross KG, Shoemaker DD, Keller L & Krieger MJB. 2005. Distribution of the two social forms of the fire ant *Solenopsis invicta* (Hymenoptera: Formicidae) in the native South American range. *Annals of the Entomological Society of America* 96, 810–817.
- Muloney S & Vanderwoude C. 2002. Red imported fire ant: an ecological threat to eastern Australia's threatened wildlife? *Ecology of Management and Restoration* 3, 167–175.
- Naurass R & Vanderwoude C. 2001. A preliminary investigation of the ecological effects of red imported fire ants (*Solenopsis invicta*) in Brisbane. *Ecological Management and Restoration* 2, 226–233.
- Poore A. 2002. Strategies for managing incursions of exotic wasps to New Zealand. *Maui* 6, 129–135.
- Parker SD & Spong DA. 1990. Invasion of polygene fire ant decimates native ants and disrupts arthropod community. *Ecology* 71, 2095–2106.
- Pomeroy SD, Williams DJ, Patterson RS & Fowler HG. 1997. Intercontinental differences in the abundance of *Solenopsis* fire ants (Hymenoptera: Formicidae): an escape from natural enemies? *Evolutionary Ecology* 11, 373–384.
- Queller DC & Goodnight KF. 1989. Estimating relatedness using genetic markers. *Evolution* 43, 256–275.
- Ross KG & Shoemaker DD. 1993. An unusual pattern of gene flow between the two social forms of the fire ant *Solenopsis invicta*. *Evolution* 47, 1595–1605.
- Ross KG, Shoemaker DD, Krieger MJB, DeHeer CJ & Keller L. 1999. Assessing genetic structure with multiple classes of molecular markers: a case study involving the introduced fire ant *Solenopsis invicta*. *Molecular Biology and Evolution* 16, 525–534.
- Schneider S, Kiechl D & Excoffier L. 2000. *Arlequin: A Software for Population Genetic Data Analysis, Ver 2.000*. Genetics and Human Laboratory, Department of Anthropology, University of Geneva, Geneva, Switzerland.
- Selby G, Vanderwoude C & Knight GK. 2002. Anophylaxis due to red imported fire ant sting. *Medical Journal of Australia* 176, 521–522.
- Silva JE & Jones RH. 2001. Top-down control by the red imported fire ant (*Solenopsis invicta*). *American Midland Naturalist* 146, 171–185.
- Tabor SV. 2000. *Fire Ants*. Texas A&M University Press, College Station, Texas, USA.
- Takezaki M & Nei M. 1996. Genetic distances and reconstruction of phylogenetic trees from microsatellite DNA. *Genetics* 144, 389–399.
- Tchinkel WR. 1998. The reproductive biology of fire ant societies. *Evolution* 48, 593–603.
- Tsutsui ND, Suarez AV, Holway DA & Case TJ. 2001. Relationships among native and introduced populations of the Argentine ant (*Linepithema humile*) and the source of introduced populations. *Molecular Biology* 10, 2151–2151.
- Vanderwoude C, Eison Harza M, Hargreaves JR, Harris E & Plowman KP. 2004. An overview of the red imported fire ant (*Solenopsis invicta*) eradication plan for Australia. *Records of the South Australian Museum Series* 7, 11–16.
- Wojcik DP, A-Jen CR, Stinner BJ, Fries EA, Jouvonen TP & Lutz RS. 2000. Red imported fire ants: impact on biodiversity. *American Entomologist* 47, 15–23.

Accepted for publication 4 February 2006